2.10 A MODEL SYSTEM FOR THE ASSESSMENT OF AMBIENT AIR QUALITY CONFORMING TO EC DIRECTIVES

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INTRODUCTION

According to the EC framework and daughter directives (*European Community, 1996, 1999, 2000, 2002*), ambient air quality must be assessed throughout the territory of member states. Concentrations shall be established at scales representative for 'at least 1000 km²', but also for 'several square kilometres' and for 'no less than 200 m²'. Limit values refer to, e.g., annual mean values or the number of exceedances for 24-hour and 1-hour average concentrations. Under certain conditions modelling techniques can be used. Since no reference techniques are specified different approaches may be considered. For the outlined purpose the model system M-SYS has been developed (Trukenmüller et al., 2004).

METHOD

The model system M-SYS (Figure 1) is based on the non-hydrostatic MEsoscale TRAnsportand Stream model METRAS (Schlünzen, 1990; Schlünzen et al., 1996), the obstacleresolving microscale model MITRAS (Schlünzen et al., 2003; Grawe et al., 2004), and the MEso- and Microscale Chemistry-Transport-Models MECTM (Müller et al., 2000, Lenz et al., 2000) and MICTM (Grawe, 2004). The two microscale models are based on the mesoscale models with respect to the equations solved, approximations and parameterisations used and the numerical schemes applied. However, the microscale models explicitly treat obstacles and are thus applicable to simulate flow fields and dispersion within the obstacle layer. All in all consistent models for the simulation of flow, transport and chemistry are used on the mesoscale and the microscale.



Figure 1. Structure of the model system M-SYS.

Three nested model areas were implemented in the mesoscale. This model hierarchy covers north-west Europe (outermost model area), the German federal state of Lower Saxony (intermediate model area) and the agglomeration Hanover-Brunswick (innermost model area). Within this innermost mesoscale model that uses a resolution of 1 km, one column (1 km by 1 km area) in the agglomeration of Hanover is simulated in the microscale model. The column is centred at Göttinger Straße in Hanover, an area that has intensely been investigated in the BMBF-funded project VALIUM by field (Schäfer et al., 2004) and wind tunnel experiments (Pascheke et al., 2002).

RESULTS

The results of the mesoscale model for the outermost model domain have been compared to DWD routine surface observations in the area of Lower Saxony. The comparison is based on the concept of hit rates (Schlünzen and Katzfey, 2003). They describe the percentage of model results within a prescribed accuracy to observations at the same place and time. The desired accuracy is taken from Cox et al. (1998) for the meteorological data. The METRAS hit rates (Table 1) are as high or even higher than the best 12-hour forecast of Cox et al. (1998). The agreement of the measured and simulated meteorological data is a precondition for a correct simulation of pollutant concentrations.

Table 1. Average hit rates for meteorological values in the M-SYS outermost model domain (grid spacing 16 km, 10 runs) and for 12-hour forecasts in a comparison by Cox et al. (1998) (grid spacing 46 km, 60 runs).

Parameter	Accuracy desired	Surface predictions within criteria		
		M-SYS	Cox et al. (best)	Cox et al. (mean)
Temperature	$\pm 2^{\circ}C$	58 %	42 %	35 %
Dew point	$\pm 2^{\circ}C$	64 %	40 %	35 %
temperature				
Wind speed	$\pm 1 \text{ m s-1}$	51 %	41 %	31 %
Wind direction	$\pm 30^{\circ}$	44 %	40 %	38 %

For chemical concentrations, accuracy requirements are defined in the daughter directives:

- For sulphur dioxide (SO₂) the maximum deviation of the measured and calculated concentration levels must not exceed 50-60 % of the hourly limit value (350 µg m⁻³),
- For nitrogen dioxide (NO_x) the maximum deviation of the measured and calculated concentration levels must not exceed 50-60 % of the hourly limit value (250 μ g m⁻³),
- For ozone (O₃) the maximum deviation of the measured and calculated concentration levels must not exceed 50 % of the hourly threshold value (180 μ g m⁻³; this value is the information threshold value).

For a total of 40 days of year 2000, concentrations modelled with M-SYS on the 16 km grid have been compared with measurements at seven NLÖ monitoring sites. All modelled concentrations of sulphur dioxide and nitrogen dioxide, giving a hit rate of 100% for this two pollutants. This fulfils the demands of the EC directives. For ozone the hit rate is slightly lower (97 %, based on the information threshold). Using the alert threshold (240 μ g m⁻³) the hit rate is again 100%. Thus, already the coarse resolution delivers quite reliable model results.

Maps for the urban background scale need a higher resolution than 16 km and thus the innermost model results are needed. For the maps with a resolution of 200 m^2 , which shall represent the air quality within the urban area, the obstacle resolving models

MITRAS/MICTM are necessary. The mesoscale models METRAS/MECTM are nested twice into the outermost model area; and profiles calculated on the innermost mesoscale model for the inner area of Hanover are eventually used to initialise the obstacle resolving microscale models MITRAS and MICTM. To ensure reliability of the flow fields, MITRAS has been evaluated in comparison with wind tunnel data (Grawe et al., 2004).

Typical results of a simulation with the microscale models MITRAS/MICTM for the area of Göttinger Straße can be seen in Fig. 2. The figure shows the concentrations of nitrogen dioxide for 11^{th} April 2003, 7.00-7:30 CET in a selected part of the model domain. The approaching flow is from the lower left corner of the figure. Emissions originate predominantly from the four traffic lines parallel to the central street canyon, but also from neighbouring main and back streets within the whole model domain ($1 \times 1 \text{ km}^2$). Figure 2a illustrates the heterogeneous distribution of the pollutants within both, the central street canyon, the neighbouring streets and the backyards.



*Figure 2. NO*₂ *concentration for 11th April, 2003, 07:00-07:30.* (*a*) *Horizontal cross section at 1.5 m above ground.*

Two distinct concentration maximums are found at the westward side of the central street canyon, while comparatively low concentration values appear on the eastward side. The distribution is strongly affected by the building structure, due to its influence on the simulated wind field. The spatial representativeness of concentration measurements within the area is very low not only for the wind field (Schlünzen et al., 2003), but also for the concentration field. Figure 2b demonstrates the three-dimensionality of the pollutant transport in the obstacle layer. Emissions originating from traffic are transported above roof level towards backyard areas.

A comparison of MICTM model results with field measurements (Schäfer et al., 2004) shows that differences are all well within a factor of 2 of the measured data (Figure 3); most differences are within 50% of the measurements. Considering the heterogeneity of the flow and concentration fields (Figure 2a), the uncertainty of the emission data and that the comparison is made for one single case only the agreement is surprisingly high. It might be derived from this case that the NO₂-fraction of 15% of the total NO_x emissions, which is used in these simulations, might be too small, since a systematic underestimation is found for NO₂ while NO is slightly overestimated from the model at all sites. At the same time NO_x values

⁽b) Vertical cross section in the area denoted by a bar in the horizontal cross section.

are quite well simulated at all sites. This suggests that the overall traffic emissions are well prescribed.



Figure 3. Comparison of simulated concentrations and field measurements within the street canyon Göttinger Straße, Hanover (11.04.03, 07:00-7:30). O3 was not measured at the selected street canyon measurement sites.

CONCLUSIONS

A multi-scale hierarchy of comprehensive models is required to reliably simulate concentrations in situations where no observations are available. Then the new approach foreseen by the directives to assess air quality from emission inventories and modelling can be applied. The model hierarchy M-SYS developed for this purpose is well applicable to simulate concentrations on different scales and with different resolutions. Comparisons with measured data show a quite good model performance, allowing to apply the model system to other regions and time periods as well. A drawback is the large amount of computer time and storage needed by the model system. Optimisation is thus necessary in order to use the model system on a routine basis. In addition, the preparation of input and comparison data needs to be simplified to reduce the amount of human power necessary for these simulations.

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