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EXPLORING THE PROPERTIES OF LOCAL AND NON-LOCAL VERTICAL DIFFUSION SCHEMES IN THE EMEP MODEL USING ²²²RN DATA

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Abstract: The simulations of hourly Radon 222 (²²²Rn) concentrations are performed with the Unified EMEP model (Simpson *et al.*, 2003) in order to validate different parameterization schemes for vertical mixing. In addition to the recently evaluated (Jeričević *et al.*, 2010) operational EMEP vertical diffusion schemes K(z), the non-local O'Brien (1970) and local Blackadar (1979) schemes, as well as the non-local Grisogono scheme (e.g. Grisogono and Oerlemans, 2002), a new scheme which is local in stable boundary layer (SBL) and non-local in convective boundary layer (CBL) and based on total turbulent energy (TTE) closure (e.g., Mauritsen *et al.*, 2007) is implemented in the EMEP model. Hourly measurements of the ²²²Rn from different stations in Europe (the Cabauw tower in the Netherlands, the Angus tower in Scotland, and Freiburg and Schauinsland in Germany) during the years 2005 and 2006 are compared to the corresponding modelled data.

Key words: higher order closure turbulence scheme, air quality modelling, Radon 222 measurements.

INTRODUCTION

The atmospheric boundary layer turbulence is the most important mechanism for the distribution of tracers. The parameterization of turbulent diffusion K(z) is an inevitable, traditional approach in description of turbulent processes and the estimation of turbulence effects in air quality models. Previous studies have already shown that the parameterizations of K(z) have significant impacts on simulated chemical concentrations (e.g. Oliviè *et al.*, 2004). Various parameterizations, mainly first-order, non-local eddy diffusivity K schemes are proposed and widely used in practical applications (e.g. O'Brien, 1970; Holtslag and Moeng, 1991; Grisogono, 1995). Several modifications of the first-order schemes are proposed to overcome deficiencies to accurately simulate dispersion in different atmospheric stability conditions (e.g. Grisogono and Oerlemans, 2002; Mihailovic and Alapaty, 2007). In this paper, we evaluate the performance of the higher-order closure K scheme based on total turbulent energy (TTE) closure (e.g., Mauritsen *et al.* 2007) in addition to recently evaluated schemes, the O'Brien, Blackadar and Grisogono schemes in the EMEP model (Jeričević *et al.* 2010). For the model evaluation available measurements ²²²Rn from Europe are used.

Radon is a radioactive gas which is found naturally in trace amounts in most rocks and soils. Since radionuclide ²²²Rn has a half-life of 3.8 days and it is emitted primarily from the continents at a fairly constant emission rate between 0.8 and 1.3 atom cm⁻²s⁻¹ (Dentener *et al.*, 1999) it is ideal to study the model sub-grid mixing schemes, numerical advection schemes or to compare different models. A considerable number of global and regional studies have been devoted to the simulation of ²²²Rn for different purposes (e.g. Lee and Larsen, 1997; Denetner *et al.*, 1999; Oliviè *et al.*, 2004; Galmarini, 2006). In this work the simulations of ²²²Rn are performed in order to validate vertical mixing schemes in the EMEP model and compare to available ²²²Rn measurements in Europe during the years 2005 and 2006. The hourly measurements of ²²²Rn from the Cabauw tower in the Netherlands, the Angus tower in Scotland, Freiburg and Schauinsland in Germany and Krakow in Poland are used. The goal of this work is to evaluate the model performance and to find the best K(z) parameterization scheme for the EMEP model as well as to better understand the behaviour of ²²²Rn in relation to the meteorological conditions.

MODEL AND METHODS

Model

The Unified EMEP model (http://www.emep.int/) was developed at the Norwegian Meteorological Institute under the EMEP programme. The model is a development of the earlier EMEP models (Berge and Jakobsen, 1998), and is fully documented in Simpson *et al.* (2003). It simulates the atmospheric transport and deposition of acidifying and eutrophying compounds, as well as photo-oxidants and particulate matter over Europe. The model domain covers Europe and the Atlantic Ocean with the grid size 50 km \times 50 km while in the vertical there are 20 terrain-following layers reaching up to 100 hPa. The Unified EMEP models uses the 3-hourly meteorological data from PARallel Limited Area Model with the Polar Stereographic map projection (PARLAM-PS), which is a dedicated version of the HIgh Resolution Limited Area Model (HIRLAM) model for use within the EMEP. In this work the Unified EMEP model version rv2_6_1 was used. In the EMEP model emissions of ²²²Rn are 1 atom cm⁻²s⁻¹ uniformly distributed over the continent.

Description of K(z) parameterization schemes

Vertical diffusion schemes, the O'Brien (1970) and Blackadar (1979) applied in convective boundary layer (CBL) and stable boundary layer (SBL) respectively, are called here the OLD K(z) scheme as they are operationally applied in the model. The OLD and Grisogono schemes (e.g. Grisogono and Oerlemans, 2002) are recently evaluated in the EMEP model (Jeričević *et al.*, 2010). Empirical coefficients determined from LES data (DATABASE64; Esau and Zilitinkevich, 2006) in stable and neutral conditions are used in the Grisogono approach (Jericevic and Vecenaj, 2009).

In this work the description of a new K(z) scheme, so called the total turbulent energy (TTE) scheme, based on a higher-order closure for neutral and stratified atmospheric conditions, is given. The TTE is the sum of the turbulent kinetic energy (E_k) and turbulent potential energy (E_p) which is proportional to the potential temperature variance. In unstable conditions the closure deploys only the TKE. Here we consider the TTE (*E*):

$$E = E_k + E_p \tag{1}$$

According to the TTE scheme vertical diffusion coefficient can be found from:

$$K(z) \approx \frac{2f_{\theta}^2 E_k l}{C_{\phi} \sqrt{E}}$$
⁽²⁾

where f_{θ} is the non-dimensional heat flux, *l* is the dissipation length scale and C_{ϕ} is the empirical constant determined based on the LES data (Mauritsen *et al.*, 2007).

Statistical methods for air quality model evaluation

It is important to properly evaluate air quality models in order to demonstrate their reliability in simulating the phenomena of interest as well as to properly test different parameterization schemes in model. Multiple performance measures are applied and considered as each measure has advantages and disadvantages and there is no single measure that is universally applicable to all conditions. In order to evaluate the predictions of a model with observations according to e.g. Wilmot (1982) and Chang and Hanna (2004) the following statistical performance measures are used in this work: the correlation coefficient (r), bias (BIAS), mean absolute error (MAE), mean square error (MSE), root mean square error (RMSE), fractional bias (FB), the normalized mean square error (NMSE), systematic ($NMSE_s$) and unsystematic ($NMSE_u$) and the index of agreement (d). The best scheme is the one which gives the best model results. The best model performance has the highest r and d, the lowest BIAS, MAE, MSE, RMSE, FB and total NMSE, while a better parameterization scheme should lower systematic errors in the model i.e. NMSEs values.

RESULTS

Measurements

In Fig. 1 normalized average monthly ²²²Rn concentrations at Freiburg and Schauinsland during 2005, at the Angus and Cabauw towers and at Krakow during 2006 are shown. The normalized average concentrations range between 0.5 Bqm⁻³ and 1.9 Bqm⁻³. The seasonal pattern is characterized by an autumn maximum and spring minimum. On average, the seasonal maximum in September is found to be higher by a factor of 3 than the April minimum. The measured concentrations are normalized due to intercomparison reasons however there is a significant difference in average values. At the Cabauw at 20m the average year concentration, $\bar{c}(^{222}Rn)$, is 1.72 Bqm⁻³, at 200 m $\bar{c}(^{222}Rn) = 1.39$ Bqm⁻³ is found, while at the Angus measured concentrations are the lowest among all analyzed stations $\bar{c}(^{222}Rn) = 0.87$ Bqm⁻³. For Schauinsland $\bar{c}(^{222}Rn) = 2.17$ Bqm⁻³ is found, while at Freiburg and Krakow concentrations are the highest among analyzed stations with $\bar{c}(^{222}Rn) = 6.27$ Bqm⁻³ and $\bar{c}(^{222}Rn) = 6.0$ Bqm⁻³ respectively.



Figure 1. Normalized average monthly ²²²Rn concentrations determined from measurements on the Angus tower in Scotland at 50 m height (blue line), the Cabauw tower in the Netherlands at 20 m (light green line) and 200 m (dark green line) heights, and in Krakow, Poland (pink line) at surface during 2006 as well as in Freiburg at 300 m (red line) and Schauinsland at 1200 m (orange line) in Germany during 2005.

The evaluation of K(z) scheme performance

The results for *FB*, *NMSE_s*, *NMSE_u* and total *NMSE* are shown in Table 1, while r and d are in Table 2. There is a significant difference in model performance at different stations. The EMEP model performs almost perfectly at the Cabauw tower with *FB* nearly equal to zero and *NMSE_s* \approx 0, while larger differences from the measurements are found at the Angus tower and Freiburg. The Grisogono scheme has the best performance at the Cabauw according to these measures. At Schauinsland the OLD scheme has slightly lower systematic error. The model has a good performance for mountain station Schauinsland. Since Schauinsland is only 8 km horizontal distance from Freiburg, and horizontal resolution in the model is 50 km x 50 km, the level closest to the height of the station is chosen as a representative for that mountain station. Results show that accuracy and systematic error in Schauinsland are low and that the chosen level is representative for the analyzed station.

It should be pointed out that the *NMSE* is reduced with the Grisogono scheme at all stations. The TTE scheme, which managed to generate the highest 222 Rn concentrations in SBL conditions, improved results at Freiburg and Krakow. Index of agreement, *d*, which is a descriptive, relative and bounded measure, as well as for *r* confirm that the best results are achieved with the Grisogono scheme at the Cabauw tower (Table 2).

Table 1. Fractional bias (*FB*), systematic part of the normalised mean square error (*NMSE_s*), unsystematic part of the normalised mean square error (*NMSE_u*) and total normalised mean square error (*NMSE*) calculated between the modelled and measured hourly 222 Rn concentrations (Bq m⁻³) for different stations: C-Cabauw tower at 200 m, the Netherlands; S-Schauinsland, Germany; K-Krakow, Poland; F-Freiburg, Germany and A-Angus tower, Scotland)

| | FB | | | NMSE_s | | | NMSE_u | | | NMSE | | |
|---------|-------|-------|-------|--------|------|------|--------|------|------|------|------|------|
| station | OLD | G | TTE | OLD | G | TTE | OLD | G | TTE | OLD | G | TTE |
| С | 0.03 | -0.09 | -0.22 | 0.00 | 0.01 | 0.05 | 0.40 | 0.33 | 0.49 | 0.40 | 0.34 | 0.54 |
| S | 0.37 | 0.39 | 0.38 | 0.14 | 0.16 | 0.15 | 0.45 | 0.51 | 0.54 | 0.60 | 0.67 | 0.69 |
| K | 0.43 | 0.42 | 0.25 | 0.20 | 0.18 | 0.06 | 1.31 | 1.34 | 1.01 | 1.50 | 1.52 | 1.07 |
| F | 0.55 | 0.52 | 0.37 | 0.32 | 0.29 | 0.14 | 0.51 | 0.51 | 0.47 | 0.83 | 0.80 | 0.61 |
| А | -0.69 | -0.64 | -0.83 | 0.54 | 0.46 | 0.83 | 0.75 | 0.48 | 1.36 | 1.29 | 0.94 | 2.19 |

Table 2. Same as Table 1 but for index of agreement (d) and correlation coefficient (r).

| | Index of agreement (d) | | | Correlation coefficient (r) | | | | |
|---------|------------------------|------|------|-----------------------------|------|------|--|--|
| station | OLD | G | TTE | OLD | G | TTE | | |
| С | 0.84 | 0.86 | 0.8 | 0.73 | 0.76 | 0.69 | | |
| S | 0.62 | 0.61 | 0.59 | 0.45 | 0.42 | 0.39 | | |
| K | 0.57 | 0.55 | 0.63 | 0.21 | 0.12 | 0.32 | | |
| F | 0.50 | 0.62 | 0.45 | 0.53 | 0.51 | 0.48 | | |
| А | 0.41 | 0.37 | 0.50 | 0.48 | 0.60 | 0.46 | | |

The model results at the Cabauw tower

In this section only measurements from the Cabauw tower are further analyzed. In Fig. 2 the time series of the observed hourly ²²²Rn concentrations are plotted against the corresponding modelled ²²²Rn concentrations calculated with three different K(z) schemes for the Cabauw tower during June, 2006. Agreement between the model and measurements is very good. The performance of the OLD and Grisogono schemes is similar while the local TTE scheme is able to capture the measured hourly peaks of concentrations ≈ 8 Bq m⁻³ and 6 Bq m⁻³, at 20 m and 200 m respectively during SBL conditions. From 1 to 14 June difference between the daily low concentrations during CBL conditions and the night-time higher concentrations during SBL conditions is obvious at 20 m, while at 200 m the daily course of concentrations is not so pronounced. This regular daily course at 20 m is interrupted in period between the 15 and 18 June due to synoptical situation. However, it should be noted that daytime mixing could be more intense to simulate the lower measured concentrations.



Figure 2. The hourly time series of the observed hourly ²²²Rn concentrations (black dots) against the corresponding modelled ²²²Rn concentrations calculated with three different K(z) schemes: the operational scheme OLD (green), the Grisogono scheme (blue) and the total turbulent energy scheme TTE (pink), at the Cabauw tower, the Netherlands during June, 2006.

In order to analyze the K(z) schemes performance separately in stable and unstable conditions two different representative cases are chosen. The modelled hourly vertical K(z) profiles and the corresponding vertical profiles of ²²²Rn concentrations are investigated. The first case is from 10 to 11 June 2006 and it is chosen from the warmer part of the year when mainly

CBL conditions prevail (Figs. 3a and 3b), and the second case is from 7 to 8 November 2006, in the colder part of the year when mainly SBL conditions prevail (Figs. 4a and 4b).

The Grisogono scheme produced lower mixing up to 100 m² s⁻¹, while much intensified mixing is produced with the OLD and TTE schemes reaching up to 400 m² s⁻¹ and 1400 m² s⁻¹ respectively during the first case in the daytime CBL conditions (Fig. 3a). On the other hand the non-local Grisogono scheme produced higher values of $K(z) \approx 6 \text{ m}^2 \text{ s}^{-1}$ in the layer near the ground of 400 m thickness during the night-time SBL conditions (Fig 3a) while the local-schemes TTE and OLD i.e. the Blackadar scheme have negligible mixing $< 0.5 \text{ m}^2 \text{ s}^1$. Note an occurrence of the intensified mixing with the TTE scheme (Fig 3a) at approximately 400 m which started to develop in the afternoon of 10th June reaching its maximum value around midnight. Obviously the TTE scheme managed to reproduce a higher turbulence in the residual layer which was not visible with the other schemes. The corresponding concentrations for the summer case with different K(z) schemes are shown in Fig 3b. A daily course in concentrations is obvious. During SBL conditions, when the mixing is low, the accumulation of the surface ²²²Rn concentrations occurs (yellow and red areas in Fig 3b). With the development of unstable conditions i.e. in CBL from 6 AM to 14 PM vertical transport is intensified, surface concentrations are diluted and higher concentrations are transported to higher levels. Neutral conditions prevail in the afternoon from 15 PM to 19 PM when the atmosphere is well mixed and the concentrations are uniformly vertically distributed. With the development of SBL nighttime conditions the accumulation starts again. Due to lower mixing in SBL concentrations produced with the TTE and OLD schemes are higher than those calculated with the Grisogono scheme. However, during CBL conditions with the Grisogono scheme concentrations are higher than with the other two schemes.



Figure 3. The modelled hourly vertical profiles for the unstable case during 10th and 11th June 2006 of a) K(z) (m² s⁻¹) and b) ²²²Rn with the OLD, Grisogono and TTE schemes, for the Cabauw tower.

The second case during 7 and 8 November 2006 is used to analyze K(z) and ²²²R profiles in the colder part of the year (Fig. 4a and 4b). The vertical mixing with all schemes is generally lower in November than in June, especially in the CBL. There is no clear difference between the night-time and day-time conditions particularly with the non-local Grisogono scheme which has $K(z) \approx 10 \text{ m}^2 \text{ s}^{-1}$ (Fig. 4a). As a result higher surface ²²²Rn concentrations are produced and mainly kept in the thin layer close to the ground (Fig 4b). Generally, the simulated surface ²²²Rn concentrations are by a factor of two higher in November than in June (Fig. 4b). During the afternoon and through the night on 8 November 2006, i.e. from 38th hour of the model run, the atmosphere was synoptically unstable due to a cold front passage over the analyzed area.

CONCLUSIONS

The evaluation of the EMEP model and K(z) parameterization schemes is based on ²²²Rn data which are found to be a good tracer to study dynamical processes in the atmosphere. Simulations of ²²²Rn with the EMEP model are performed during the years 2005 and 2006 and compared to the available ²²²Rn measurements in Europe: the Cabauw and Angus towers, Freiburg, Schauinsland and Krakow. In addition to recently evaluated the OLD and Grisogono schemes (Jeričević *et al.*, 2010), a new scheme which is based on total turbulent energy (TTE) closure (Mauritsen *et al.*, 2007) is implemented in the EMEP model and analyzed. Intercomparison of different local and non-local schemes on the ²²²Rn data showed that the non-local scheme Grisogono scheme (Jeričević and Večenaj, 2009) are primarily developed for neutral and stable conditions based on LES data (DATABASE64; Esau and Zilitinkevich, 2006). The estimation of empirical coefficients for CBL conditions on the LES data is foreseen. The local schemes produce higher surface concentrations in SBL conditions, while the Grisogono scheme K(z) is determined independently at each model level based on local vertical gradients. The present version of the EMEP model has the lowest level at 100 m which is an important deficiency for non-local K(z) schemes to properly simulate diffusion in SBL conditions. In order to evaluate the model predictions with observations and to estimate the performance of different K(z) schemes a set of statistical measures is used (e.g. Wilmot, 1982; Chang and Hanna, 2004).

Results of the model evaluation on ²²²Rn data showed that the model has the best results for the Cabauw tower. The Cabauw tower is representative for the model evaluation due to its position in a flat terrain as well as due to uniform ²²²Rn emission in the area. On the other hand data in Freiburg and Krakow are affected by the local natural emissions of ²²²Rn while the Angus tower concentrations are dominated by the advection of ²²²Rn free air from the sea (the emissions of ²²²Rn are 100 times less

over the sea than over the land). The highest concentrations are simulated with the TTE scheme and systematic error is decreased while accuracy in increased in the model for Freiburg and Krakow. For an appropriate description of ²²²Rn distribution in the atmosphere, its response to the latitudinal, time and intensity variability of precipitation should be accounted explicitly (Galmarini, 2006). Since those variabilities in radon natural fluxes are not included in the model certain deviations from observations are expected.



Figure 4. The modelled hourly vertical profiles for the stable case during 7^{th} and 8^{th} November 2006 of a) K(z) (m² s⁻¹) and b) ²²²Rn with the OLD, Grisogono and TTE schemes, for the Cabauw tower.

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REFERENCES

- Blackadar, A. K., 1979: Modeling pollutant transfer during daytime convection. In: Fourth Symposium on Atmospheric Turbulence Diffusion and Air Quality, AMS, Reno, NV, pp. 443-447.
- Berge, E., and Jakobsen, H. A., 1998: A regional scale multi-layer model for the calculation of long-term transport and deposition of air pollution in Europe, *Tellus*, **50**, 205–223.
- Chang, J.C., and Hanna, S.R., 2004: Air quality model performance evaluation, Meteorol. Atmos. Phys., 87, 167-196.
- Dentener, F., Feichter, J., and Jeuken, A., 1999: Simulation of the transport of 222Rn using on-line and off-line global models at different horizontal resolutions: a detailed comparison with measurements, *Tellus B*, **51**, 573-602.
- Esau, I., and Zilitinkevich, S., 2006: Universal dependences between turbulent and mean flow parameters in stably and neutrally stratified planetary boundary layers, *Nonlinear Proc. Geoph.* **13**, 135-144.
- Galmarini, S.: One year of 222-Rn concentration in the atmospheric surface layer, Atmos. Chem. Phys. 6, 2865–2887, 2006.
- Grisogono, B.: A generalized Ekman layer profile within gradually-varying eddy diffusivities, *Q. J. Roy. Meteorol. Soc.*, **121**, 445-453, 1995.
- Grisogono, B. and J. Oerlemans, 2002: Justifying the WKB approximation in pure katabatic flows, Tellus A, 54, 453-462.
- Holtslag, A. A. M., and Moeng, C. H., 1991: Eddy diffusivity and countergradient transport in the convective atmospheric boundary layer, *J. Atmos. Sci.*, **48**, 1690–1698.
- Jeričević A., and Večenaj, Ž., 2009: Improvement of vertical diffusion analytic schemes under stable atmospheric conditions, *Boundary-Layer Meteorol.*, **131**, 293-307.
- Jeričević, A., L. Kraljević, B. Grisogono, H. Fagerli, and Ž.Večenaj, 2010: Parameterization of vertical diffusion and the atmospheric boundary layer height determination in the EMEP model, *Atmos. Chem. Phys.*, **10**, 341-364, 2010.
- Lee, H. N., and Larsen, R. J., 1997: Vertical diffusion in the lower atmosphere using aircraft measurements of 222Rn, J. *Appl. Meteorol.*, **36**, 1262-1270.
- Mauritsen, T., Svensson, G., Zilitinkevich, S., Esau, I., Enger, L., and Grisogono, B., 2007: A total turbulent energy closure model for neutrally and stably stratified atmospheric boundary layers, *J. Atmos. Sci.*, **64**, 4113-4126.
- Mihailovic D T, Alapaty K, 2007: Intercomparison of two K-schemes: local versus nonlocal in calculating concentrations of pollutants in chemical and air-quality models. *Environ Modell Soft*, **22**, 1685 1689.
- O'Brien, J. J., 1970: A Note on the vertical structure of the eddy exchange coefficient in the planetary boundary layer, J. *Atmos. Sci.*, **27**, 1213-1215.
- Oliviè, D. J. L., Van Velthoven, P. F. J., and Beljaars, A. C. M., 2004: Evaluation of archived and off-line diagnosed vertical diffusion coefficients from ERA-40 with 222Rn simulations, *Atmos. Chem. Phys.*, **4**, 2313-2336.
- Simpson, D., H. Fagerli, J. E. Jonson, S. Tsyro, P. Wind and J.-P. Tuovinen, 2003: The EMEP Unified Eulerian Model. Model Description. Technical Report EMEP MSC-W Report 1/2003, The Norwegian Meteorological Institute, Oslo, Norway.
- Willmott, C.J., 1982: Some comments on the evaluation of model performance, Bull. Am. Meteorol. Soc., 63, 1309-1313.